

## Zooplankton (Cladocera, Copepoda) dynamics in the River Danube upstream and downstream of Budapest, Hungary

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**Abstract.** The spatial distribution and seasonal dynamics of zooplankton (Cladocera, Copepoda) were studied in the River Danube near Budapest, Hungary. The investigated river section was relatively poor in plankton, nauplii dominated. A total of 36 species was recorded of which *Acanthocyclops robustus*, *Thermocyclops crassus*, *Bosmina longirostris* were the most abundant. There was a downstream increase in copepod densities, however, no other remarkable differences could be observed between the profiles upstream and downstream of the capital. Generally, the streamline was characterized by lower densities and lower number of taxa as compared to the river bank; nevertheless, there were differences between the left and the right banks both upstream or downstream as well. Seasonal dynamics was defined by a marked late winter–spring aspect and abundance peaks were found to be characteristic for Danube with high densities in May–June and August–September. Examining the relationship between zooplankton density and the hydrological regime, it can be concluded that zooplankton production in the main channel is of minor importance, rather floodplain areas and adjacent water bodies seem to be important sources of plankton biomass.

River Danube, as being the most species rich river in the Palaearctic (Naidenow, 1998), is entitled to great interest, so the research of its plankton means a crucial challenge. The first investigations on the zooplankton of the Hungarian section of the River Danube started in the 1910s with crustaceans (Kottász, 1913; Jungmayer, 1914; Unger, 1916) and continued in the middle of the century (Woynárovich, 1944; Éber, 1955; Ponyi, 1962) mainly with descriptive works. Nevertheless, these studies served as a basis for further research. Dudich (1967) gave detailed information about the fauna of the river Danube. Bothár has broken new ground in the studies of riverine plankton since the 1960s. With research among others on the sampling strategies (Bothár, 1996), on the long-term status of planktonic crustaceans at the area of Göd (Bothár, 1968, 1972, 1988a, 1994, 1996) and on the crustacean plankton near Budapest (Bothár, 1978, 1988b), she has made a major contribution to our better understanding of zooplankton in the River Danube.

In the 1970s and particularly in the 1980s crustacean abundance increased notably, which was best professed by the abundance peaks in summer. The change in species composition and individual numbers implied eutrophication in Danube. How-

ever, the hydrological regime has changed as well, which may influence zooplankton assemblages both qualitatively and quantitatively (Bothár, 1985). Towards the estuary, crustacean density increased by 25 fold within the 417 km long Hungarian river section (Bothár, 1988 a). The third period of the Danube plankton research began in the 1990s with the studies of Gulyás (1994, 1995, 1997, 2002) who has extended the research to rotifers. Rotifers dominated the plankton, only copepodites and nauplii were represented in similar abundance.

The most frequent species were characteristic of eutrophicated stagnant waters and rivers of low current velocity (Gulyás, 1994, 1995, 2002). Large abundance of some species and the downstream increase of zooplankton abundance implied increasing eutrophication in the river (Gulyás, 1995). Biomass values measured upstream and downstream of Budapest (near Budapest) were similar, then increased southwards (Gulyás, 1997).

Although numerous studies have been conducted on the zooplankton in the River Danube, relatively little emphasis was put on the transversal distribution of plankton (Bothár, 1978, 1985; Naidenow, 1971, 1979). On the other hand, de-

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tailed surveys have been performed on the horizontal distribution of zooplankton along the river (Bothár, 1973, 1982, 1988 a; Gulyás, 2002; Pujin 1990; Naidenow & Schewzowa, 1990, 1991).

The objectives of this study were (1) to get a comprehensive picture of the crustacean plankton near Budapest, (2) to describe the community structure both from qualitative and quantitative points of view, (3) to present the spatial and temporal changes of microcrustacean plankton considering the hydrological regime as a possible driving force.

## MATERIALS AND METHODS

### Study sites

River Danube is the second longest river in Europe, it is more than 2800 km long with a catchment area of 817,000 km<sup>2</sup>. The Hungarian section occupies 417 km. The middle section is 377 km long and has a relatively balanced hydrological regime, which is controlled by the water discharge of the upper section. In the middle section, current velocity ranges between 0.8–1.2 m sec<sup>-1</sup> in the upper layers, which can reach the values of 2–2.5 m sec<sup>-1</sup> during flooding events. The water level fluctuation takes 5–8 m, discharge values range from 500 (low water period) to 6000–8000 (flood) m<sup>3</sup> s<sup>-1</sup>. The shoreline is mostly regulated, rip-rap is characteristic of this section.

The present study was conducted in the main channel of the River Danube at two profiles upstream and downstream of Budapest (Újpest–Békásmegyér, 1657 rkm and Tököl–Százhalombatta, 1623 rkm) (Fig. 1). The general characteristics of the two profiles are similar, but the sampling site of Tököl–Százhalombatta is characterized by gravel banks interrupted occasionally with rip-rap, whereas in the profile of Újpest–Békásmegyér rip-rap is the main component. There was no macrovegetation either at the upstream or at the downstream profile throughout the years. The sampling point of Békásmegyér is situated in the downstream section of the Danube arm of Szentendre, near to the estuary into the main channel.

The study sites were designated on the basis of our objectives, i.e. possibilities for transversal sampling (ferry), and sampling profiles upstream and downstream of Budapest.

### Sampling and data analysis

Samples were collected at biweekly intervals from October 2006 to September 2008 at two sampling profiles each containing three sampling sites (stream line, left bank and right bank). Samples taken at the river bank were performed at water depth of some 1–2 metres. During the winter period (between December and February) zooplankton was sampled monthly. 100 litres of water were taken from the surface water layer (upper 50 cm water column) and filtered through a plankton net (50 µm mesh size). The material collected was preserved *in situ* in 4% formaldehyde solution. A total of 270 samples were collected and analysed. Nauplii were counted in 5 ml subsamples in special counting chambers after homogenization. For the taxonomic determination of the animals identification keys by Gulyás and Forró (1999, 2001), Einsle (1993), Amoros (1984) and Dussart (1969) were used. Copepods and cladocerans were identified to species level, however, copepods belonging to the suborder Harpacticoida and ostracods were only counted.

Water temperature was measured *in situ* whereas conductivity was measured in laboratory. Water discharge and water level data (1645.5 rkm) were obtained from the Environmental and Water Research Institute („VITUKI”). Water residence time was calculated with the formula:

$$R = 0.08A_d^{0.6}/Q^{0.1}$$

where R is the residence time at the sampling site (days), A<sub>d</sub> watershed area upstream of the sampling site (km<sup>2</sup>), and Q river discharge (m<sup>3</sup> s<sup>-1</sup>) (Soballe & Kimmel, 1987). We used the discharge data measured at Budapest (1646.5 rkm). In order to explore the temporal patterns, cluster analysis and non-metric multidimensional scaling (NMDS) using the Euclidean distance were performed (with standardized data).

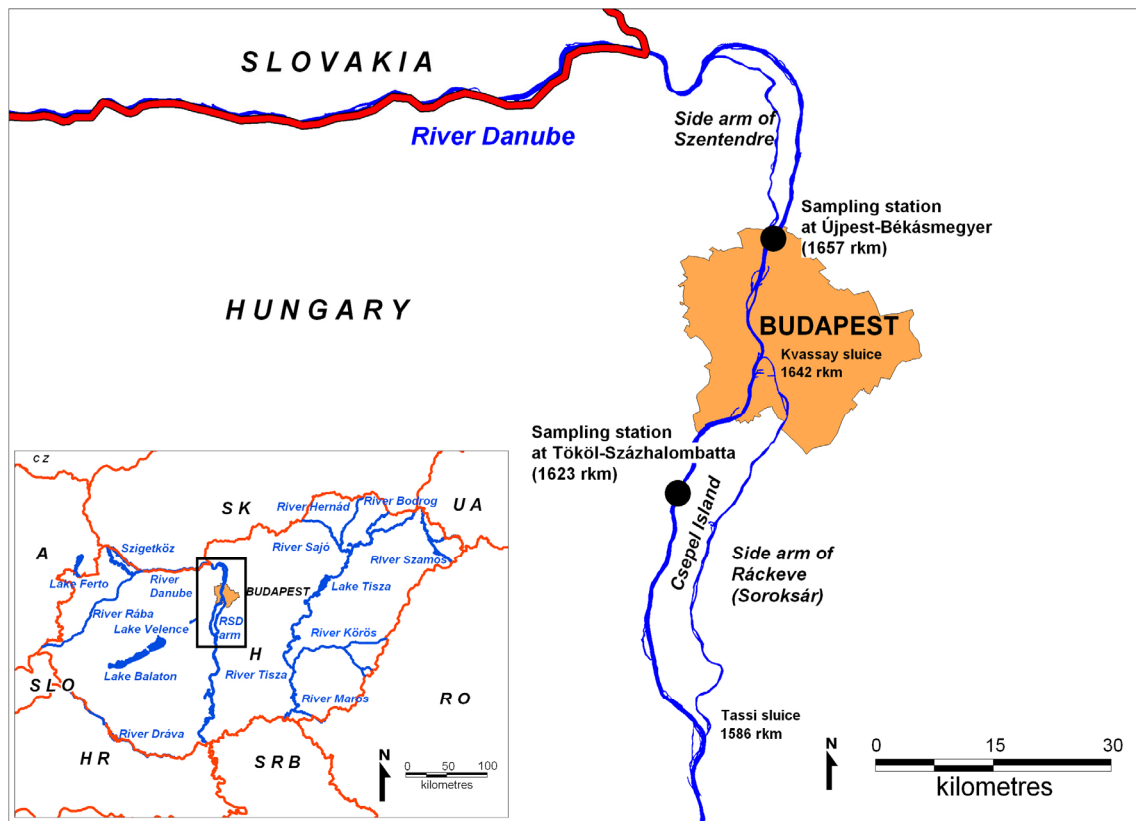
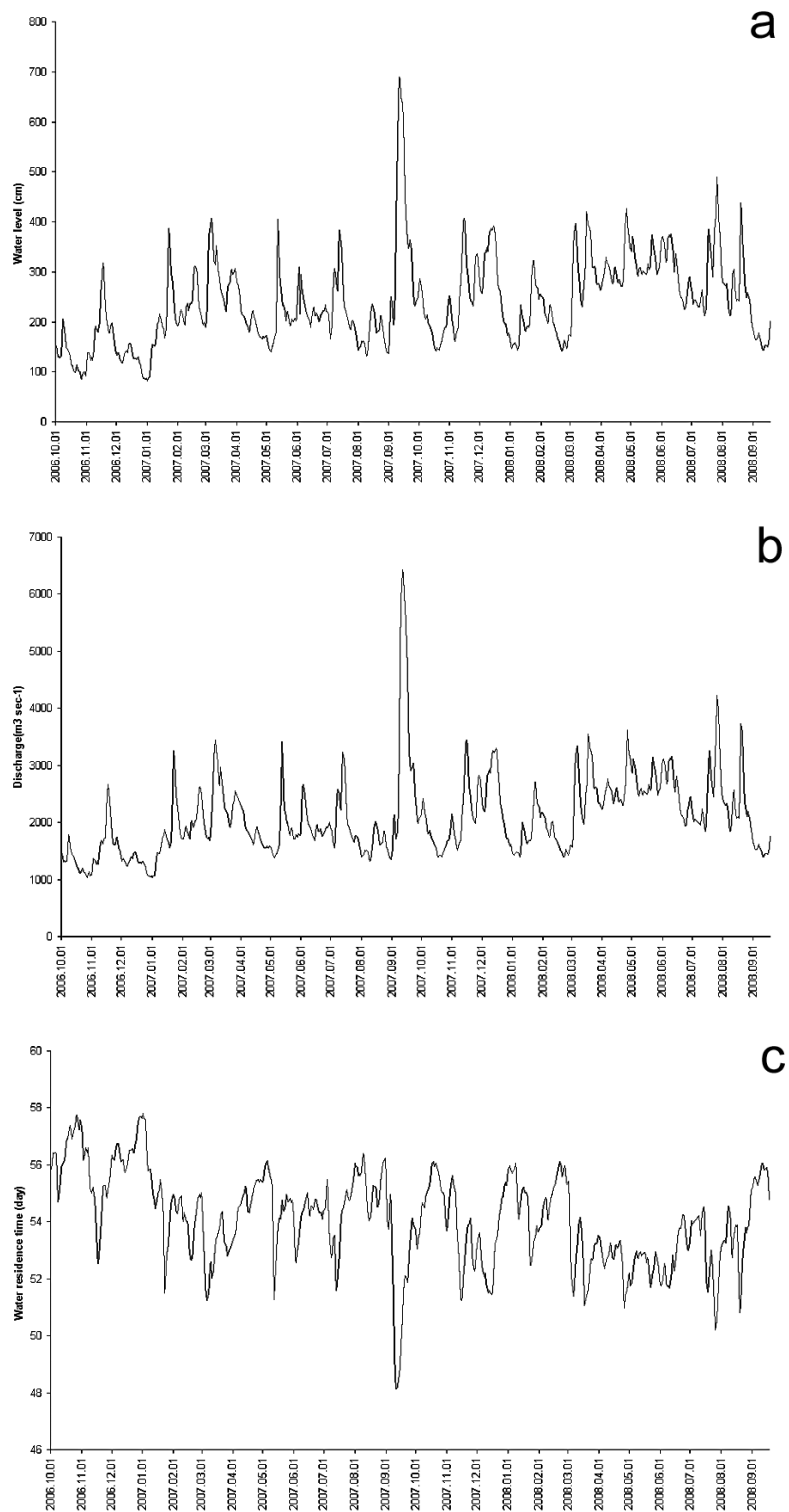


Figure 1. The study area and the sampling sites (●)

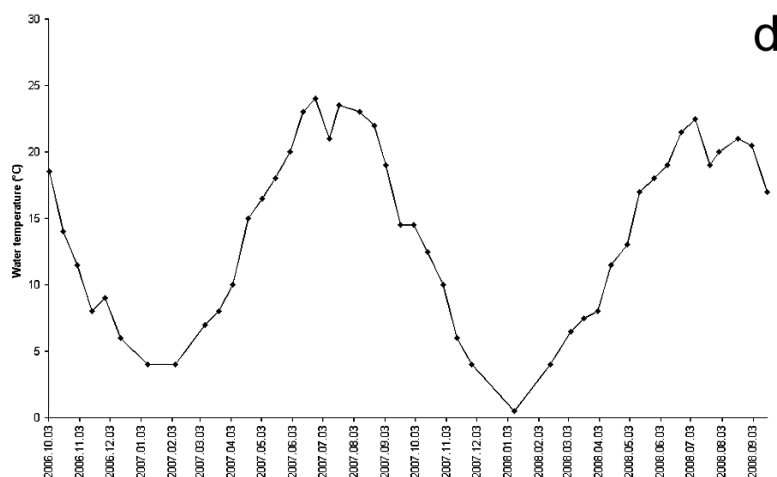
Correspondence analysis was applied to demonstrate the relationship between species and sampling sites, that is certain species seem to be more abundant at a definite sampling site. Diversity t-test is a useful tool to compare the diversities of the sampling sites by calculating the Shannon diversity of each sampling site and comparing the diversities statistically. Linear correlation was used to detect any significant association between some environmental variables and zooplankton community. When necessary, data were transformed (log, square root) to obtain the normal distribution. Significant differences were identified at  $p < 0.05$ . All data analyses were performed using the PAST program (Hammer *et al.*, 2001).

## RESULTS

During the study period, water level varied between 83 and 689 cm-s, maxima were recorded in September of 2007. This marked peak was two times higher than other peaks. Low water level occurred in the year of 2006, whereas relatively high water level persisted in spring and summer of 2008. Discharge values (ranged between 1030 and 6420  $\text{m}^3 \text{sec}^{-1}$ ) strongly followed this pattern. Despite some deviations, the annual discharge pattern typical of river systems in the temperate region was observed with low water period in autumn and winter, and discharge peaks in early and late spring. Water residence time ranged between 48 and 58 days (Figs. 2 a–d). Water



Figures 2 a-c. Hydrological parameters recorded during the study period. a = water level, b = discharge, c = water residence time



**Figure 2 d.** Water temperature recorded during the study period

temperatures showed typical annual cycle, maximum values were observed in June–August (23–24°C) and minimum in January (0.5 °C). There was a negative relationship between water level and conductivity both upstream ( $r = 0.40$ ;  $p = 0.0073$ ) and downstream ( $r = 0.38$ ;  $p = 0.0118$ ) of Budapest over the study period.

A total of 36 species were observed of which 11 belonged to copepods and 25 to cladocerans (Table 1). *Diacyclops crassicaudis* was recorded for the first time in the Hungarian section of the River Danube. The occurrence of *Diaphanosoma mongolianum* in the main channel of the Danube is new, as it was only reported from the Szigetköz backwaters and few water bodies respectively (Gulyás & Forró, 1999). Frequent species included *Acanthocyclops robustus*, *Thermocyclops crassus*, *Bosmina longirostris*, *Alona rectangula*, *Chydorus sphaericus*, moreover harpacticoid copepods were also determining. The species number in the streamline was similar upstream and downstream (17–16 species), 10 common species were recorded. The distribution of the species number within the sampling profile was more dispersed downstream with a maximum of 30 species detected at Tököl. The relative abundance of nauplii was mostly greater than that of other zooplankton groups, whereas adult copepods and copepodites were represented in similar abundances. Clado-

cerans contributed up to 19.5% of the total density upstream of Budapest and 18.1% downstream, respectively.

The percentage compositions of the zooplankton community upstream and downstream of the capital are presented in Figs. 3–4. There is evidence of a remarkable downstream increase in copepod densities meaning that adult copepod densities became doubled and copepodite densities increased by 18%. This phenomenon is mainly due to the fact that the relative contribution of *Thermocyclops crassus* increased considerably at the downstream profile, although calanoid densities increased as well. Such notable longitudinal differences cannot be seen either in cladoceran community composition, or in nauplii.

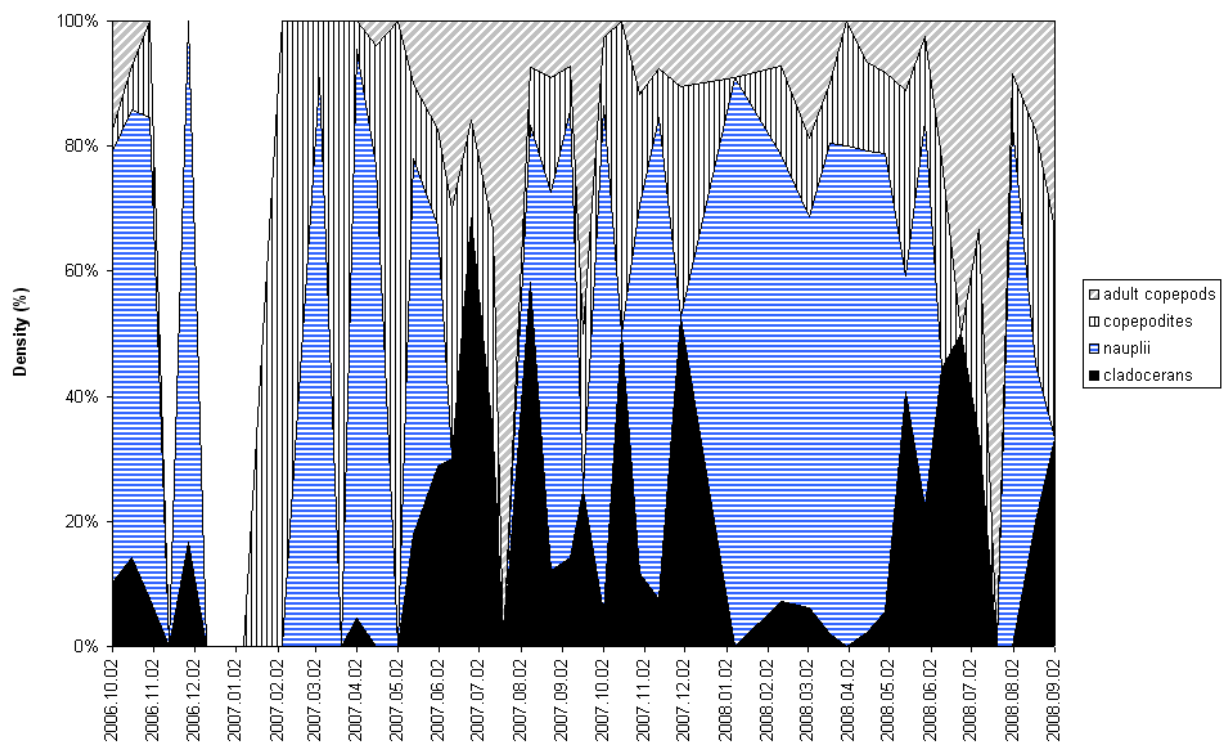
Generally, the streamline was characterized by lower individual numbers and lower number of taxa in comparison with the river bank (Table 2), however, there were also differences between the left and the right banks both upstream or downstream of Budapest. Average densities ranged between 5.08 and 12.96 ind./100 litres, whereas maximum density reached 87 ind./100 litres at Békásmegyér. Note that nauplii densities are included as well, these results suggest that the river is poor in planktonic crustaceans. Regarding diversities it seemed to be essential to present the

**Table 1.** Zooplankton taxa recorded in the River Danube (October 2006 to September 2008) and their codes. Abbreviations: B: Békásmegyer (right bank); UBS: Újpest–Békásmegyer streamline; U: Újpest (left bank); Sz: Százhalombatta (right bank); TSzS: Tököl–Százhalombatta streamline; T: Tököl (left bank)

Taxa	Code	B			UBS			U			Sz			TSzS			T		
		06	07	08	06	07	08	06	07	08	06	07	08	06	07	08	06	07	08
<b>CLADOCERA</b>																			
<i>Alona affinis</i> (Leydig, 1860)	Aaff		+						+								+		+
<i>Alona guttata</i> Sars, 1862	Agut		+																+
<i>Alona intermedia</i> Sars, 1862	Aint		+																
<i>Alona quadrangularis</i> (O. F. Müller, 1785)	Aqua	+	+			+			+	+		+	+						+
<i>Alona rectangula</i> Sars, 1862	Arec	+	+	+	+	+			+			+	+	+			+		+
<i>Bosmina coregoni</i> Baird, 1857	Bcor																		+
<i>Bosmina longirostris</i> (O. F. Müller, 1785)	Blon		+	+	+	+	+		+	+		+	+		+	+	+	+	+
<i>Chydorus sphaericus</i> (O. F. Müller, 1776)	Csph	+		+		+	+	+	+	+			+		+	+			+
<i>Daphnia cucullata</i> Sars, 1862	Dcuc		+			+	+		+	+		+	+		+	+	+	+	+
<i>Daphnia longispina</i> O. F. Müller, 1785	Dlon		+			+			+						+				+
<i>Diaphanosoma brachyurum</i> (Liévin, 1848)	Dbra		+						+										
<i>Diaphanosoma mongolianum</i> Uéno, 1938	Dmon												+						+
<i>Disparalona rostrata</i> (Koch, 1841)	Dros						+		+	+			+						+
<i>Graptoleberis testudinaria</i> (Fischer, 1848)	Gtes		+																
<i>Iliocryptus sordidus</i> (Liévin, 1848)	Isor			+					+			+	+		+				+
<i>Leydigia acanthocercoides</i> (Fischer, 1854)	Laca																		+
<i>Leydigia leydigi</i> (Schoedler, 1863)	Lley								+			+							+
<i>Macrothrix hirsuticornis</i> Norman & Brady, 1867	Mhir			+					+	+		+	+			+			+
<i>Macrothrix laticornis</i> (Fischer, 1848)	Mlat								+										+
<i>Moina macrocopa</i> (Straus, 1820)	Mmac		+						+										
<i>Moina micrura</i> Kurz, 1874	Mmic		+			+	+		+			+	+						+
<i>Pleuroxus aduncus</i> (Jurine, 1820)	Padu											+							+
<i>Scapholeberis mucronata</i> (O. F. Müller, 1785)	Smuc		+				+		+							+			
<i>Sida crystallina</i> (O. F. Müller, 1776)	Scry																		+
<i>Simocephalus vetulus</i> (O. F. Müller, 1776)	Svet		+																
<b>COPEPODA</b>																			
<b>CALANOIDA</b>																			
<i>Eudiaptomus gracilis</i> (Sars, 1863)	Egrac						+						+						+
<i>Eurytemora velox</i> (Lilljeborg, 1853)	Evel		+						+	+		+	+		+	+			+
<b>CYCLOPOIDA</b>																			
<i>Acanthocyclops robustus</i> (Sars, 1863)	Arob		+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+
<i>Acanthocyclops vernalis</i> (Fischer, 1853)	Aver															+			
<i>Cyclops strenuus</i> Fischer, 1851	Cstr									+	+	+							+
<i>Cyclops vicinus</i> Uljanin, 1875	Cvic			+			+		+										+
<i>Diacyclops crassicaudis</i> (Sars 1863)	Dera						+												
<i>Eucyclops serrulatus</i> (Fischer, 1851)	Eser		+	+		+	+		+			+	+			+			+
<i>Mesocyclops leuckarti</i> (Claus, 1857)	Mleu		+																+
<i>Paracyclops fimbriatus</i> (Fischer, 1853)	Pfim		+				+		+	+									+
<i>Thermocyclops crassus</i> (Fischer, 1853)	Tera		+	+		+	+		+	+	+	+	+		+	+			+
<b>HARPACTICOIDA</b>																			
<i>Har</i>	Har	+	+	+		+	+	+	+	+	+	+	+		+	+	+	+	+
<b>OSTRACODA</b>																			
<i>Ost</i>	Ost		+						+	+						+			+

**Table 2.** Major ecological parameters of the zooplankton community (n = 44). Density data include nauplii as well. For the abbreviations of the sampling sites see Table 1

Sampling site	B	UBS	U	Sz	TSzS	T
Average density (ind./100L)	8,48	6,80	9,82	9,17	5,08	12,96
Maximum density (ind./100L)	87,00	49,00	43,00	46,00	22,00	60,00
Taxa S	26,00	17,00	25,00	19,00	16,00	30,00
Dominance D	0,23	0,13	0,08	0,11	0,13	0,14
Shannon H	2,21	2,31	2,83	2,52	2,32	2,48
Evenness e <sup>H</sup> /S	0,35	0,59	0,68	0,66	0,64	0,40
Berger-Parker	0,45	0,26	0,16	0,19	0,21	0,25



**Figure 3.** The percentage composition of the zooplankton community upstream Budapest. Results include data obtained from the streamline, left and right river banks (summed)

results of two different diversity indices. Shannon diversity values ranged within a relatively close interval (2.21–2.83), highest diversity and evenness values were measured at Újpest. Berger–Parker diversity peaked at Békásmegyér where the Shannon diversity was the lowest, which is due to the dominance of some species (dominance  $D =$

0.23). Based on the diversity t-test, sampling station Újpest (U) differed significantly from the others (U-B:  $t = 3.89$ ,  $p < 0.001$ , U-UBS:  $t = 3.77$ ,  $p < 0.001$ , U-Sz:  $t = 2.29$ ,  $p = 0.02318$ , U-TSzS:  $t = 3.68$ ,  $p < 0.001$ , U-T:  $t = 2.54$ ,  $p = 0.0156$ ). One more significant difference included Békásmegyér–Százhalombatta ( $t = 2.32$ ,  $p = 0.0214$ ).

The correspondence analysis (Fig. 5) suggests that the sampling site Békásmegyer, which is situated in the side arm of Szentendre, has specific features. The high individual numbers of *Moina micrura* and the presence of *Simocephalus vetulus*, *Alona guttata*, *Mesocyclops leuckarti*, *Graptoleberis testudinaria* are responsible for its separation. The sampling site Újpest (U) is also determined by the presence of several taxa. The sampling profile downstream of Budapest seems to form one group, although the sampling site Tököl (T) has some characteristic elements, e. g. *Bosmina coregoni*, *Sida crystallina* and it is dominated by *Thermocyclops crassus*.

Examining the temporal patterns of the zooplankton assemblage, no clear trajectory could be detected, however, there was a marked late winter-spring aspect (marked) with great similarities among samples. Samples taken in summer and autumn form a second group, but are far more dispersed (Fig. 6). Species number was low during the winter and increased in late spring, then remained relatively high during the summer months and decreased in autumn. Total abundance, however, peaked in spring of 2008 and in summer of 2007. The seasonal dynamics can be characterized by two population peaks within the year, the first in May–June, the second in August–September.

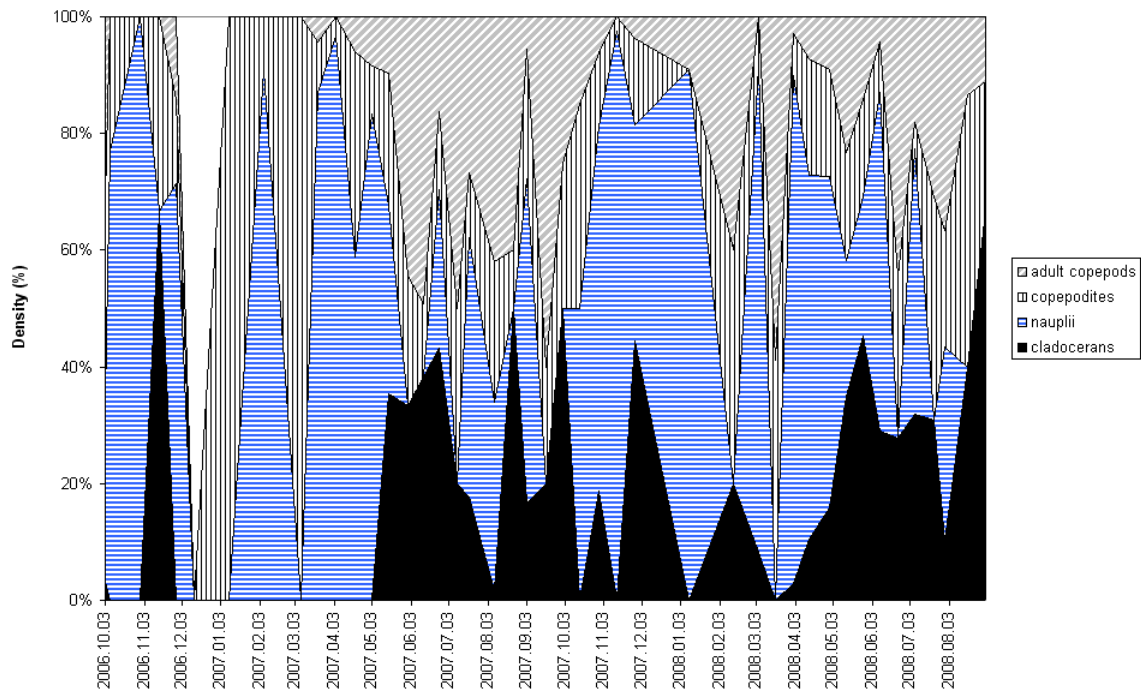
The seasonal patterns were similar between the upstream and downstream sections, with a positive correlation of the zooplankton community density between the two sampling profiles over the study period ( $r = 0.58$ ;  $p < 0.001$ ). A strong positive correlation was detected between water temperature and zooplankton density at Tököl-Százhalombatta ( $r = 0.35$ ;  $p = 0.0178$ ), but not at Újpest-Békásmegyer, however, zooplankton density without nauplii was positively related to water temperature ( $r = 0.42$ ;  $p = 0.004$ ). Zooplankton density (without nauplii) was positively correlated with water level ( $r = 0.32$ ;  $p = 0.0303$ ) and discharge ( $r = 0.32$ ;  $p = 0.0307$ ) and negatively correlated with water residence time ( $r = 0.32$ ;  $p = 0.0305$ ) at the upstream section. However, these relationships were not significant at the downstream profile.

## DISCUSSION

The time available for zooplankton to develop was estimated at 48–58 days (an estimate of the time the water has been in the river system) assuming that plankton drifts with the current passively. During this time 5 generations of cladocerans and 1 or 2 generations of copepods may develop (Naidenow, 1998). The growth is supposed to depend on the hydrological regime (increased residence time favours zooplankton, that is zooplankton benefits from low water velocity). This is not only due to the mechanical effect of the drift, but it is connected with the fact, that suspended matters have a negative impact on planktonic crustaceans (Zsuga *et al.*, 2004; Gulyás, 2002;). However, we have found a negative relationship between water residence time and zooplankton density and a positive relationship between water discharge and zooplankton density, which strongly suggests that zooplankton production in the main channel is of minor importance, which is consistent with the findings of Reckendorfer *et al.* (1999). Floodplain areas and adjacent water bodies seem to be rather important sources of plankton biomass (Saunders & Lewis, 1989; Naidenow 1998; Schiemer *et al.* 2001; Zsuga *et al.*, 2004). The positive correlation of zooplankton density with water level also seems to support this hypothesis. According to Reckendorfer *et al.* (1999), the physical interaction of flow regime and river margin morphology determines the availability of inshore storage zones and the rate at which plankton are added to the main river channel.

The dominant species that were found in the present study in the main channel (*Acanthocyclops robustus*, *Bosmina longirostris*) are consistent with the findings of Bothár (1985, 1988 b, 1994) and Gulyás (1994, 1995, 2002), however, the relatively large contributions of *Thermocyclops crassus* and Harpacticoida are new. Two thermophil species, *Thermocyclops crassus* and *Moina micrura* are reported to become abundant in the River Danube since 1971 (Bothár, 1975), but they were regarded only as secondary species at Göd (rkm 1669) (Bothár, 1985). The relatively





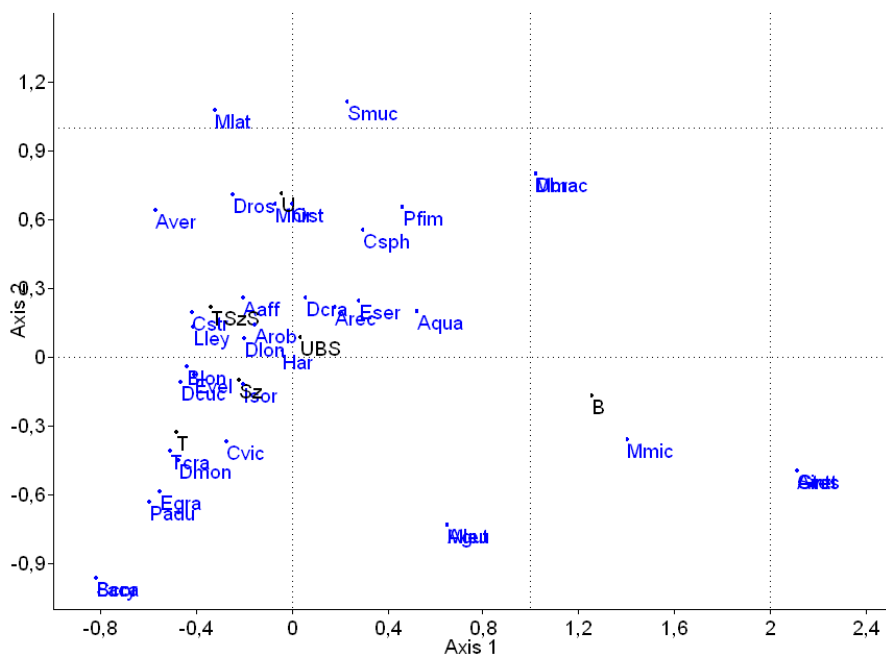
**Figure 4.** The percentage composition of the zooplankton community downstream Budapest. Results include data obtained from the streamline, left and right river banks (summed)

large frequency and abundance of *Thermocyclops crassus* may be related to the changing hydrological regime and increased temperature. Beyond the above-mentioned, there is no evidence of any significant long-term changes in species composition. The population peaks observed in the present study (May–June, August–September) are in line with the results of Bothár (1985, 1978), Bothár and Kiss (1990) and Gulyás (1995). Regarding densities, the examined river stretch was poor in plankton, the individual numbers were similar to the findings of Bothár *et al.* (1971) and Bothár (1978) at Újpest–Békásmegyer in the 1970s, at Adony (1598 rkm) in the 1980s (Bothár, 1988b), but not to that of Göd (Bothár, 1994). In quantitative aspect, contrasting my results to that of Gulyás (1994, 1995, 1997) is not meaningful since the author presented the overall densities (with rotifers).

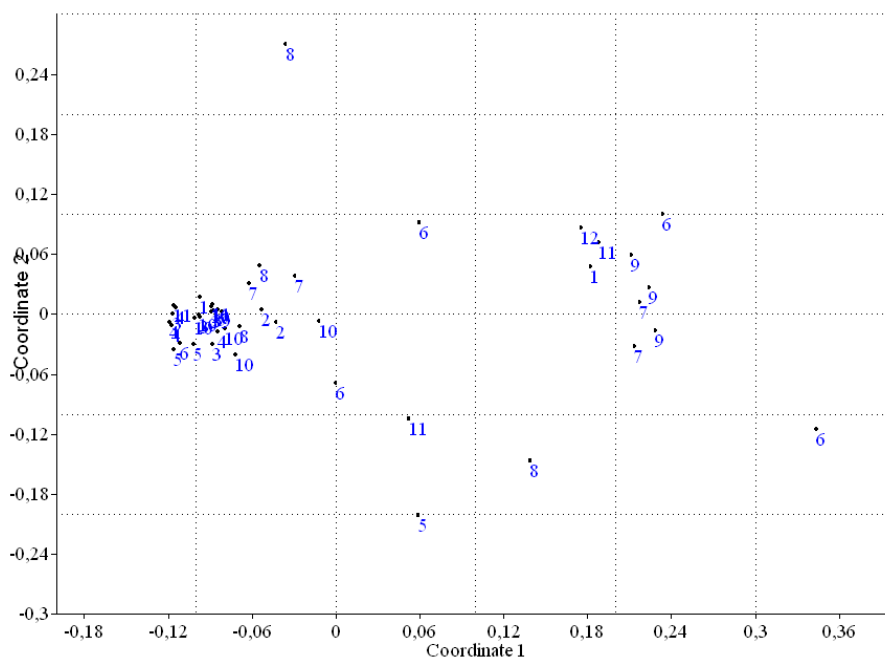
There were no remarkable differences between the upstream and downstream sections, only copepod density increased downstream. This does

not support the hypothesis that municipal and industrial wastewaters of the capital may create significant impact on zooplankton assemblages downstream Budapest. Bothár (1988) has observed lower maximum densities and different seasonal dynamics patterns downstream of Budapest. According to the author, these differences can be explained by the effects of wastewaters and changed stream conditions due to the regulated river bed and the side arm of Szentendre. Nevertheless, Gulyás (1997) has not found any remarkable differences between the upstream and downstream sections. Although the examined river stretch is poor in planktonic crustaceans, the recorded species and their contributions to the total density suggest that the river has a moderately rich fauna with several tichoplanktonic elements. The latter also implies for the significance of inshore storage zones.

The spatial distribution of zooplankton across the river was not equal, the streamline was characterized with lower densities and lower number



**Figure 5.** Biplot of the correspondence analysis. Codes of the taxa are presented in Table 1. Abbreviations of the sampling sites: B-Békásmegyér; UBS-Újpest-Békásmegyér streamline; U-Újpest; Sz-Százhalombatta; TSzS-Tököl-Százhalombatta streamline; T-Tököl



**Figure 6.** NMDS plot of the samples. Numbers represent the months: 1-January; 2-February etc.

of taxa. Similarly, densities were often found to be higher nearshore (Bothár, 1978, 1985; Thorp *et al.*, 1994; Mitsuka & Henry, 2002).

In summary, my results pointed out the evidence of spatial heterogeneity across the river, however, save adult copepods no major difference was detected between the upstream and downstream sampling profiles. The author stresses the importance of adjacent lentic areas as sources of planktonic crustaceans.

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